Mercury contamination in East Fork Poplar Creek, Oak Ridge, Tennessee

G.R. Southworth Environmental Sciences Division Oak Ridge National Laboratory



Mercury issues in Oak Ridge

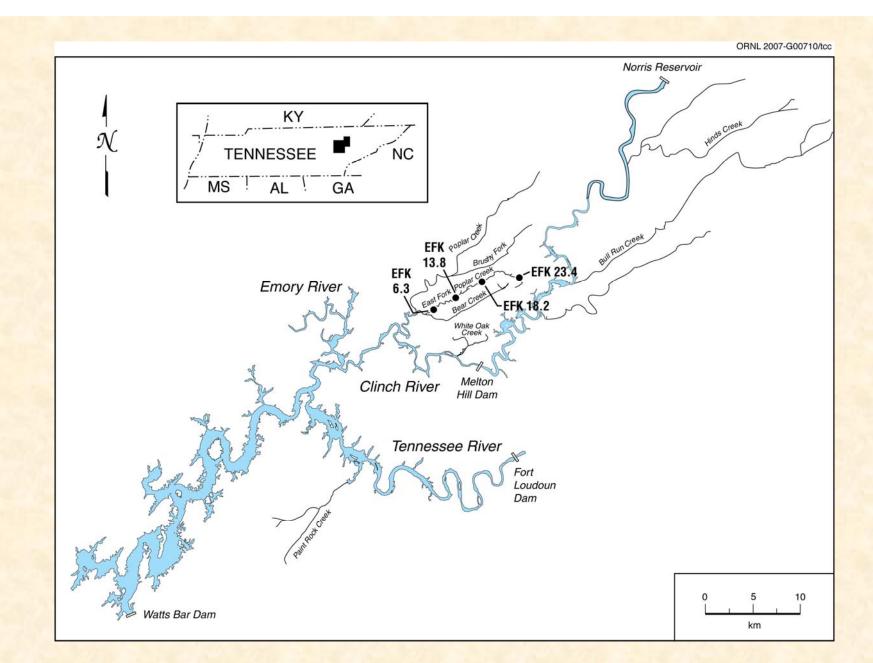
Industrial use of metallic mercury in 1950's and 1960's contaminated soil, buildings, storm drain network, ground and surface water.



1.1 million kilograms of mercury were lost at the site, with about 10% of that going to East Fork Poplar Creek.

Processes using mercury were discontinued in 1963.











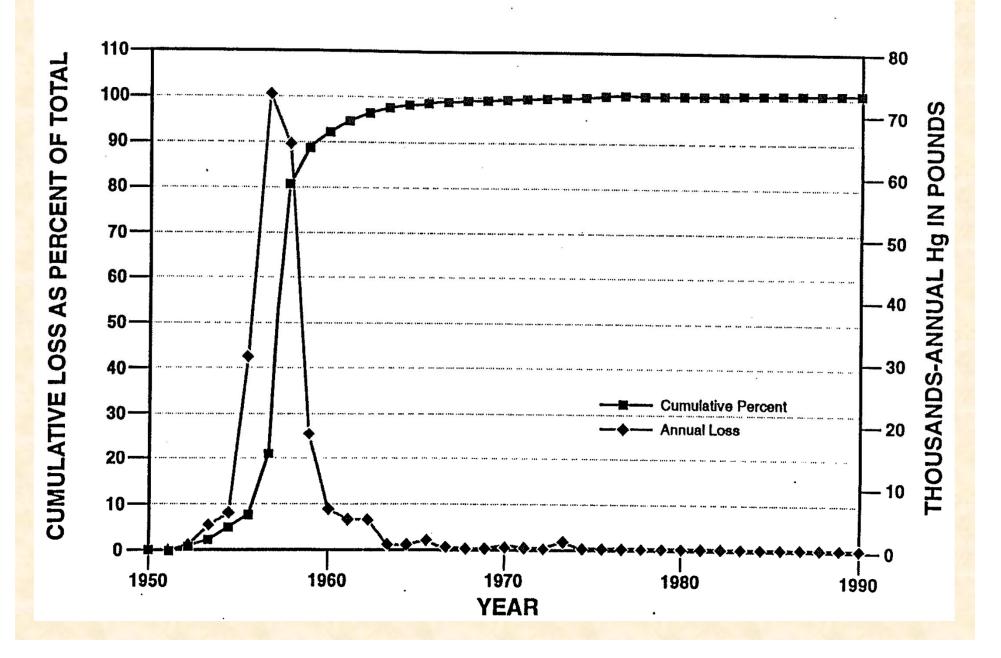
History of mercury actions in EFPC



- 1950s-1960s: Industrial use of metallic mercury at Y-12 resulted in contaminated soil, buildings, storm drain network, ground and surface waters.
 - Approximately 2.4 million pounds of mercury were lost at the site, with about 10% of that going to East Fork Poplar Creek.
- 1963: Processes using mercury at Y-12 were discontinued.
- 1988 New Hope Pond replaced
- 1990's RMPE EEMTS & CMTS construction, storm drain cleanout / lining
- 1992 Dechlorination of discharge water
- 1996 Flow Management established base flow
- 1998 Lake Reality bypassed
- 2001 Bank stabilization in UEFPC to limit Hg soil erosion
- 2005 Big Spring Water Treatment System



Mercury Losses to EFPC



East Fork Poplar Creek setting progresses from industrial to urban to agricultural to woodland



Remedial actions have focused on controlling methylmercury bioaccumulation by reducing the concentration of waterborne inorganic mercury





Success of that approach depends upon MeHg bioaccumulation being limited by the concentration of inorganic Hg



Sources of mercury to EFPC

- Mercury use area, storm drain network
- Metallic mercury in streambed sediments
- Metallic mercury in solution cavity network (karst system)
- Erosion of Hg-contaminated streambank soils and streambed sediments
- Erosion of Hg-contaminated soils (floodplain, scrapyard, etc)
- Background mercury (rain, uncontaminated soils)



Source: Storm drain network above N/S pipe, historic Hg-use area

~ 3 – 10 g/d loading

Unique chemistry, Hg solubilized by HOCI

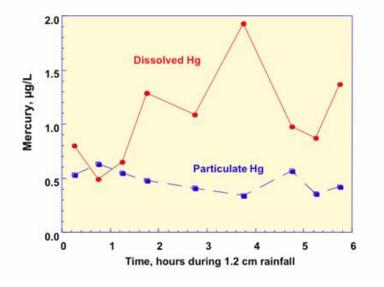


Metallic mercury in gravelfilled pipe

OAK RIDGE NATIONAL LABORATORY U. S. DEPARTMENT OF ENERGY

Response to rain suggests dissolved Hg sources close to pipes

Mercury in storm drain discharge versus time during rain



10

JT-BATTELLE

Source: Metallic mercury in streambed sediments

Blobs of Hg metal on clay hardpan under armored soft sediments Generates >30 µg/L Hg Input to surface flow enhanced by Flow Management

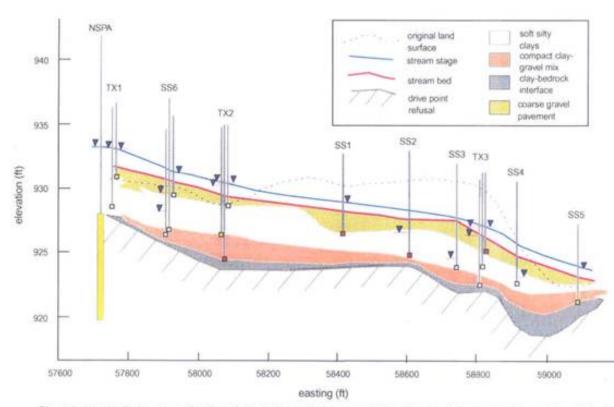




Figure 4. Longitudinal cross section through the study reach showing water levels, subsurface geometerials, and Hg distribution.

Upper East Fork Poplar Creek Longitudinal Stream Transect

Effect of Flow Management on mercury flux and concentration

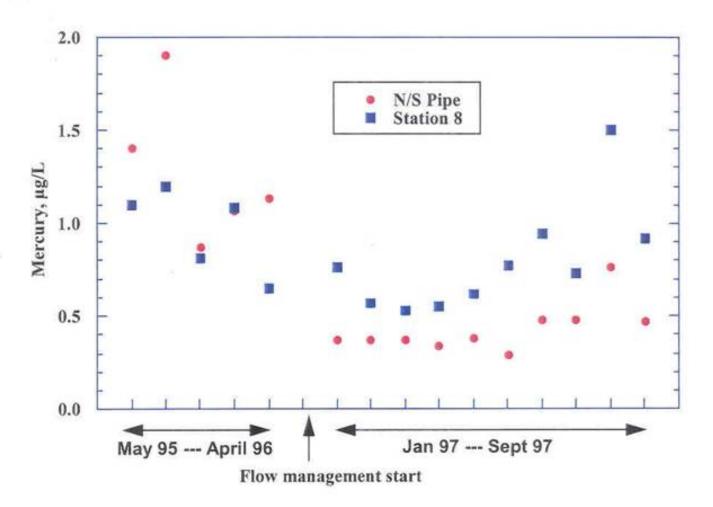


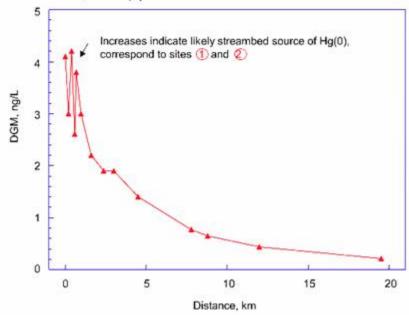
Fig. 1. Total mercury concentrations in UEFPC below the N/S Pipe and at Station 8, before and after Flow Management. Post-Flow Management samples were taken after mixing with raw water.



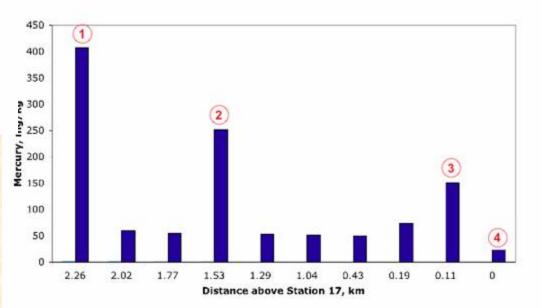
OAK J U.S.D

Indications of streambed mercury

Dissolved gaseous mercury (Hg(0)) versus distance from North/South pipe at Y-12



Mercury in fine sediment suggests metallic Hg in streambed at sites 1 and 2



Outfall 51, a Hg-contaminated spring, contributed about 3 - 4 g Hg/day to EFPC, most of which was highly reactive dissolved Hg(II) and Hg(0)



Tracer dye added to karst system 800 m upstream emerges from spring Oak Ridge National Laboratory U. S. DEPARTMENT OF ENERGY Activated charcoal treatment removes > 99% of Hg





Source: Mercury eroded from streambank and streambed

Primary source of wet weather loading Dissolved Hg doesn't increase during wet weather loading Flow Management raised water level to contact highly contaminated 'black layer'



Eroding 'black layer' contained average 300 ppm Hg, highest at waterline (max 2213 ppm)

After



Fig. A.6. UEFPC Bank Stabilization final site with riprap and geotextile tubes filled with sand.

Fig. A.1. UEFPC Bank Stabilization site before construction.

Before bank stabilization

Wet weather loading

Response to rainfall event indicates that the short (1,000 m) reach of open stream from N/S pipe to Station 8 is a much greater source of Hg loading during rain than the entire watershed above the N/S pipe.

That pattern continued in next 1,000 m.

OAK RIDGE NATIONAL LABORATORY U. S. DEPARTMENT OF ENERGY

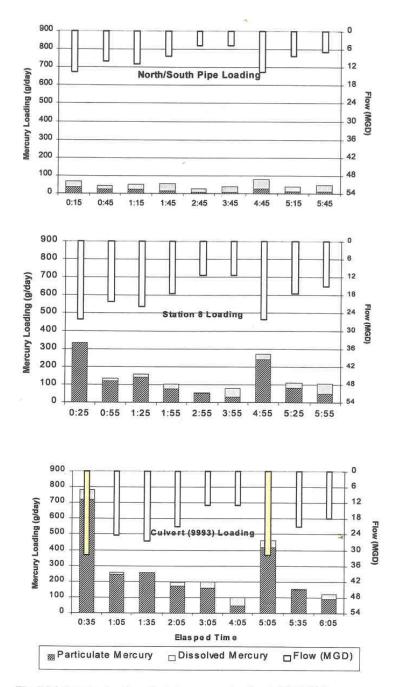


Fig. 3.15. Dissolved and particulate mercury loading in UEFPC during a storm on January 7, 1998.

Source: Erosion of mercury-contaminated soils

Lower EFPC floodplain Soils > 400 ppm Hg removed under CERCLA Estimated loading to EFPC in 1984 ~ 500 g/d, but almost all associated with wet weather transport

Effects on baseflow Hg transport hard to discern

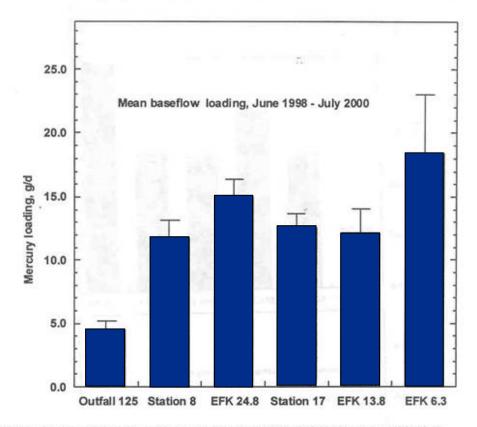
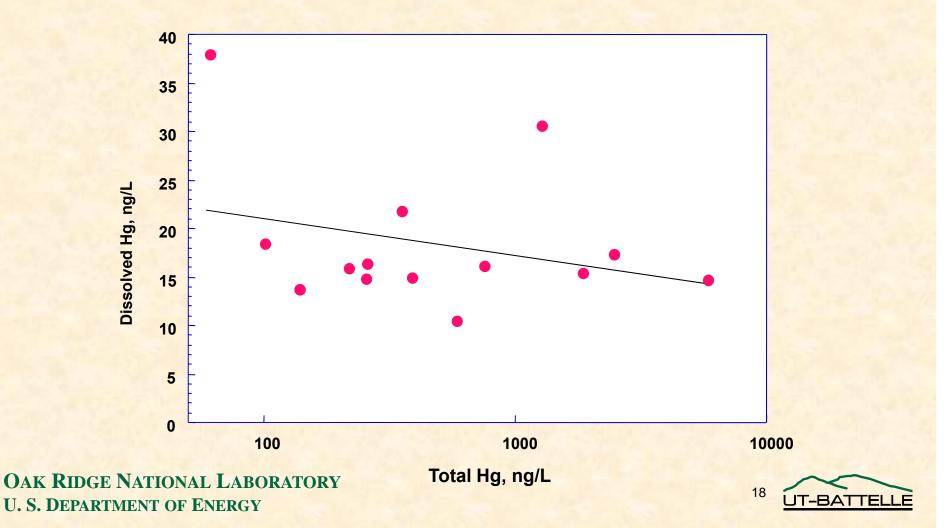


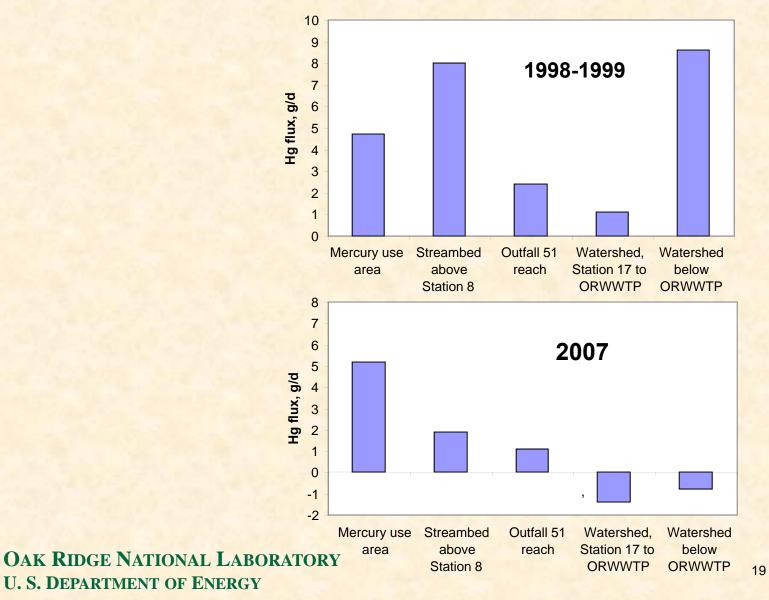
Fig. 4.4. Downstream changes in mercury loading in EFPC. Values are mean ± SE; N=8.



Dissolved Hg concentrations decrease during high flow when total Hg concentrations and Hg loading are highest

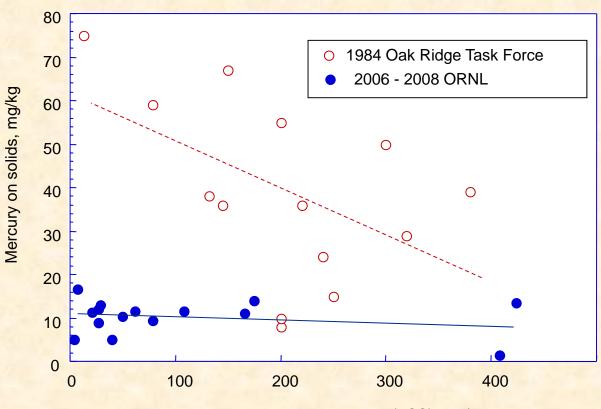


Baseflow mercury loading by source (N=5), 1998-1999 and 2007 (N=1)





Wet weather flux 1984 versus 2007-2008

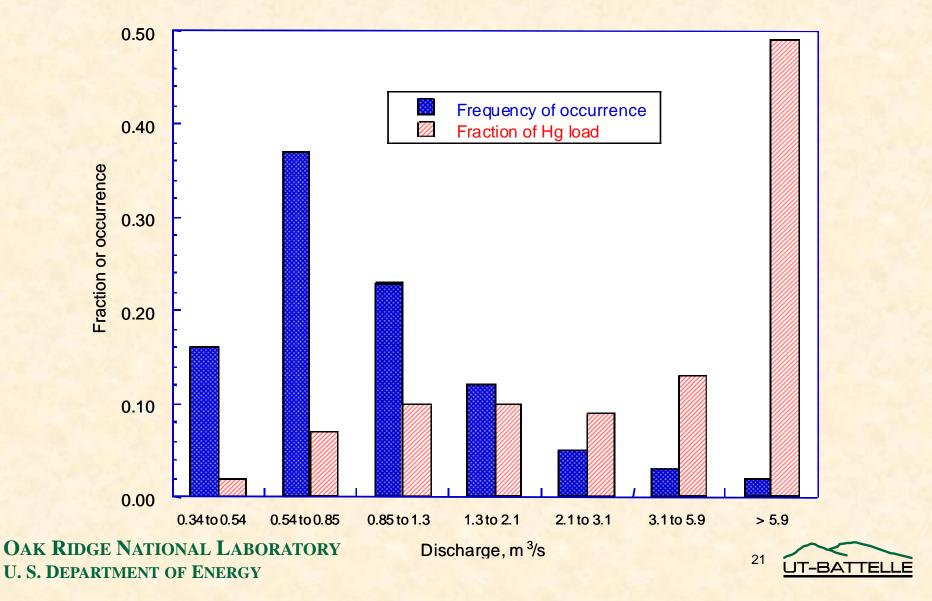


Total suspended solids (TSS), mg/L

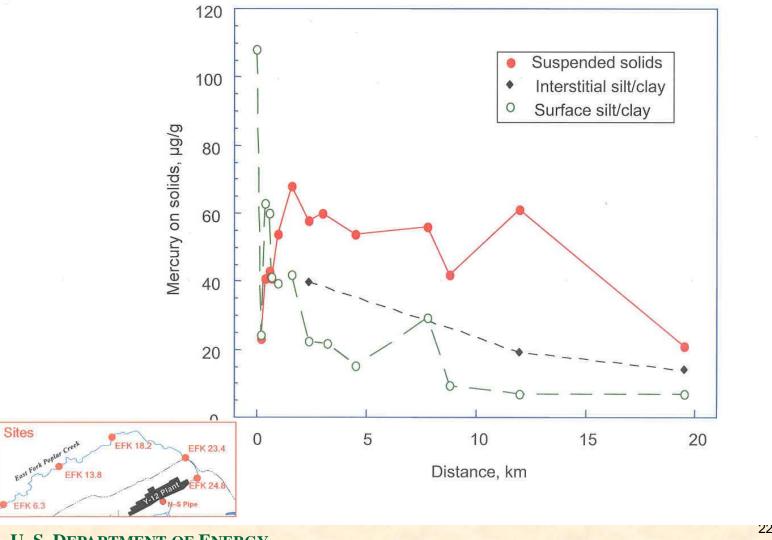
ORTF 1984 flux from watershed below Y-12 - 193 kg/y 2007/2008 estimated flux % reduction - 73% (52 kg/y)



Most Hg transport was predicted to occur in infrequent events



Concentrations of Hg on particulates (surface sediment, interstitial silt, and suspended solids versus distance from the N/S pipe at Y-12

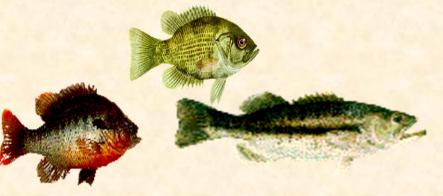


22 UT-BATTELLE

U. S. DEPARTMENT OF ENERGY

Bioaccumulation monitoring approach in EFPC

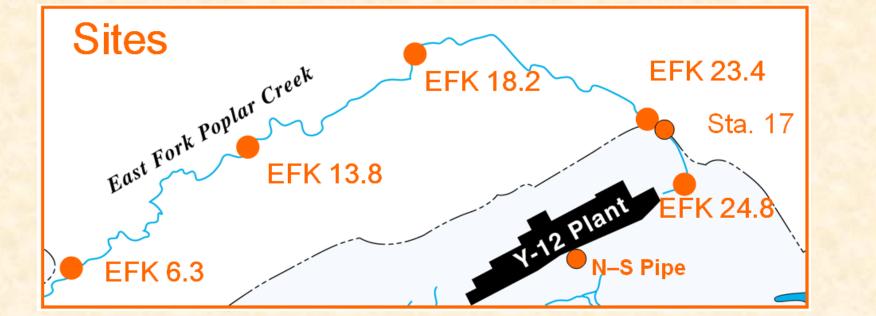
- Monitoring of resident sunfish primarily (redbreast, rockbass)
- Five sites throughout length of 25 km stream
- Twice yearly sampling
- 6-8 individual fish fillets/site
- Edible sized fish targeted, similar sizes between sites and years







Hg in Fish Tissue : Spatial and Temporal Trends

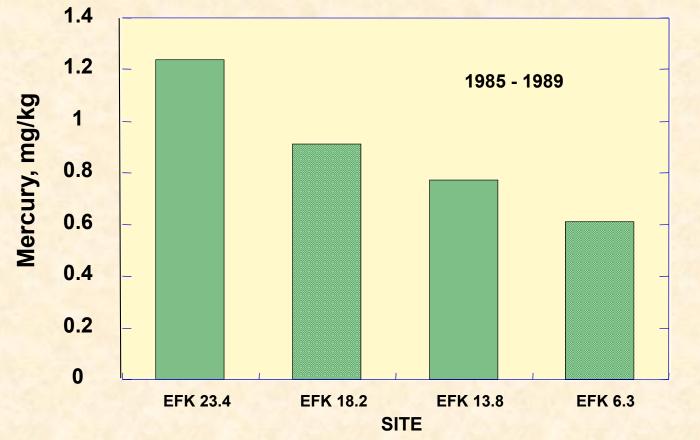


Facility Abatement Actions, 1984-present

EVENT:			Pond replaced						Dechlorination				Flow management			passed		Bank stabilized		
				\leftrightarrow				\leftrightarrow				\leftrightarrow			\leftrightarrow			\leftrightarrow		
YEAR:	1984	1985	1986	1987	1988	1989	1990	1991	1992	1993	1994	1995	1996	1997	1998	1999	2000	2001	2002	2003

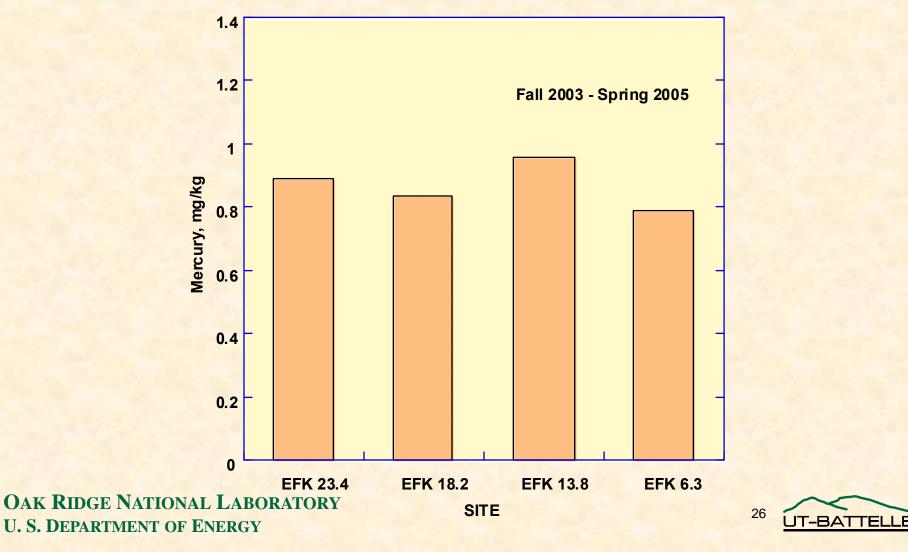


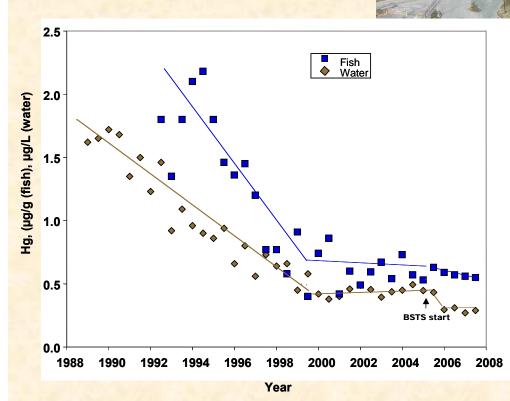
The downstream profile of Hg in fish in EFPC in 1980's was consistent with downstream dilution of a headwater point source. Headwater Hg loading > 100 g/d Total residual chlorine ~ 0.5 ppm in upper 2 km.





Since the early 1990's, the downstream profile of mercury in fish has been uniform throughout the creek Change coincides with dechlorination of all process water discharges



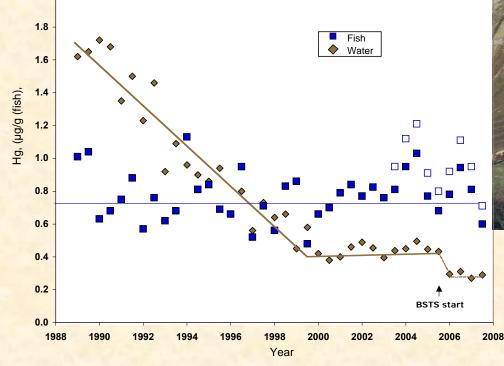


OAK RIDGE NATIONAL LABORATORY U. S. DEPARTMENT OF ENERGY



Hg in fish and water have changed commensurately in upper EFPC (above the pond in the photo)





Upper East Fork Poplar

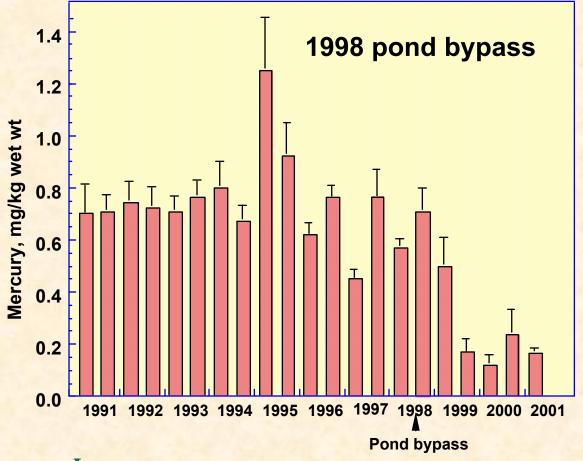
Hg in fish has not responded to upstream decrease in Hg in water at site below the pond

OAK RIDGE NATIONAL LABORATORY U. S. DEPARTMENT OF ENERGY

2.0



Eliminating waterborne Hg (but not 50 - 100 ppm Hg in sediments) produced a striking decrease in Hg bioaccumulation in bluegill in the pond

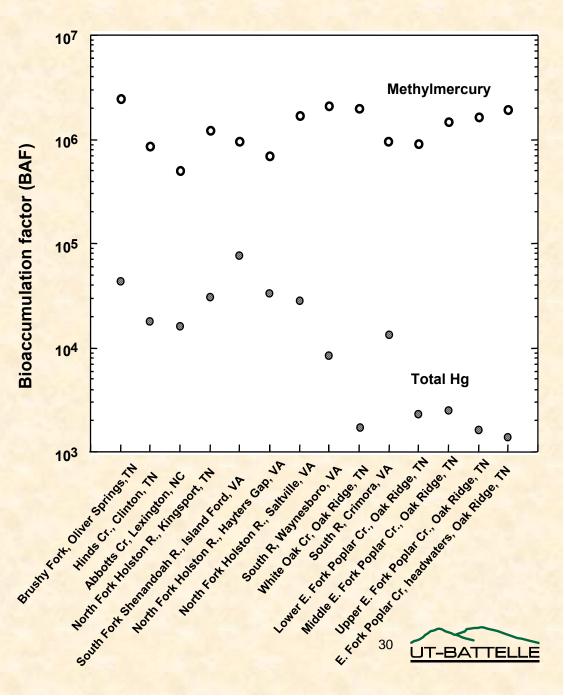


OAK RIDGE NATIONAL LABORATORY U. S. DEPARTMENT OF ENERGY

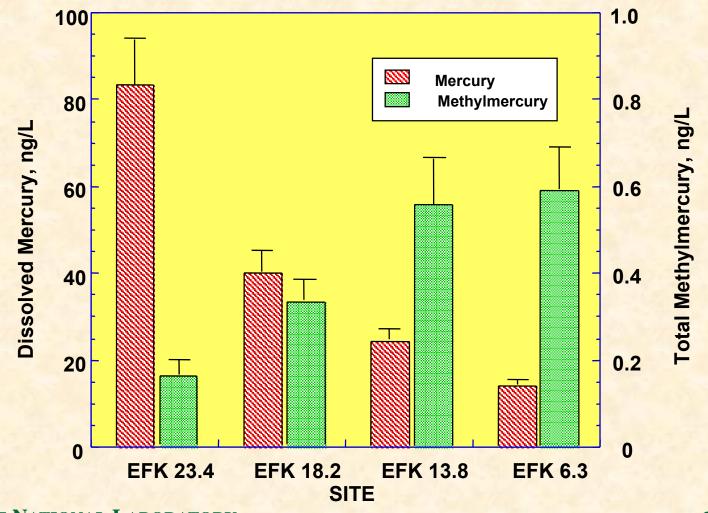
29 UT-BATTELLE

Bioavailability

- Bioavailability of Me-Hg in EFPC appears similar to uptake factors in other streams
- Bioavailability of Hg-Total in EFPC is lower than other sites (contaminated and uncontaminated)



Dissolved Hg does not appear to be directly related to methylmercury production

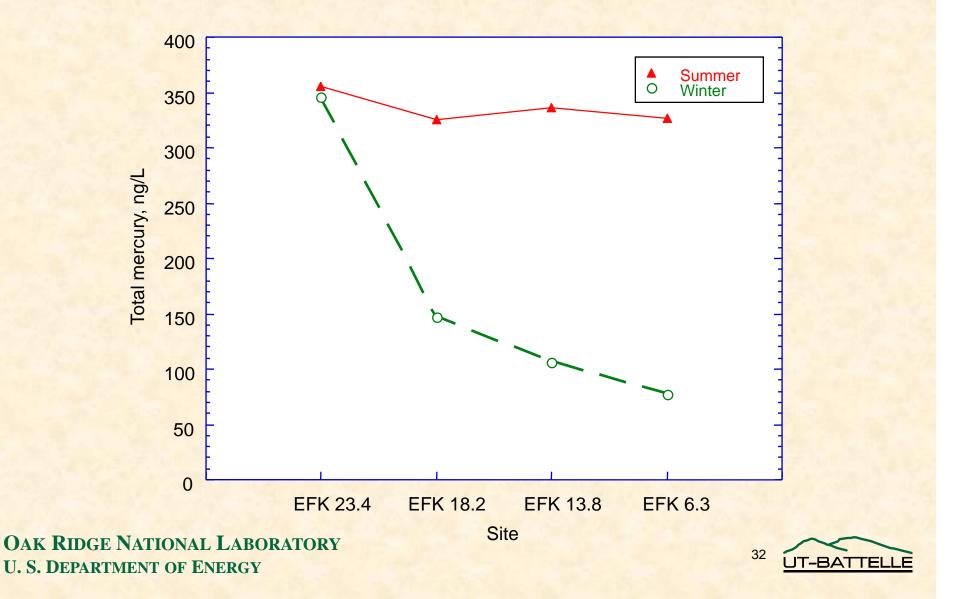


OAK RIDGE NATIONAL LABORATORY U. S. DEPARTMENT OF ENERGY

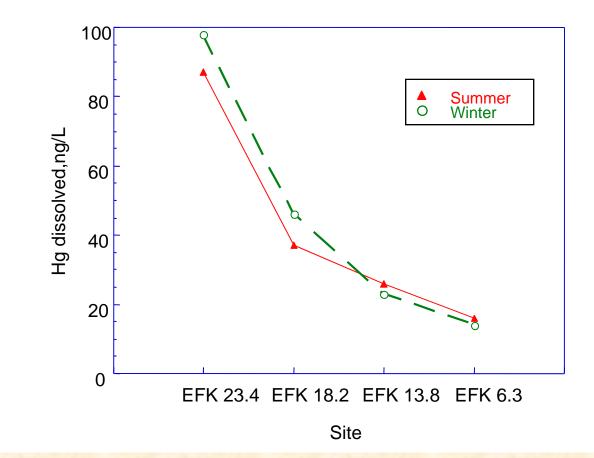
31 UT-BATTELLE

Downstream profiles of total mercury concentration in EFPC,

2000 - 2006, winter versus summer

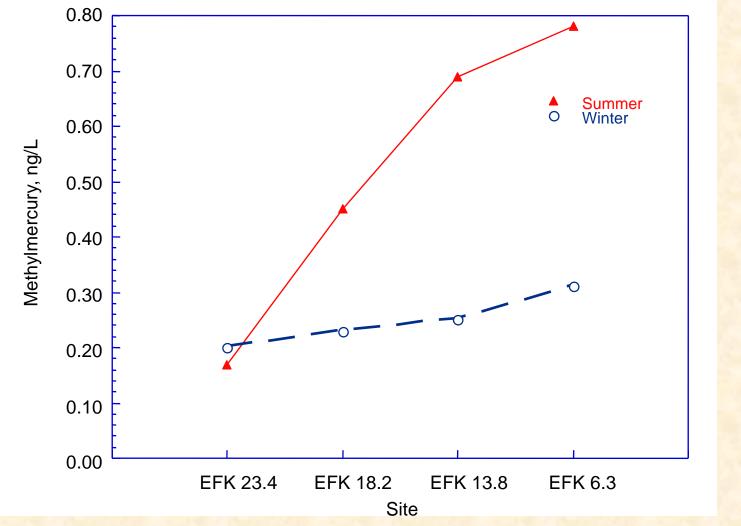


Downstream profiles of dissolved total mercury concentrations (filtered) in EFPC, 2000 - 2006, winter versus summer



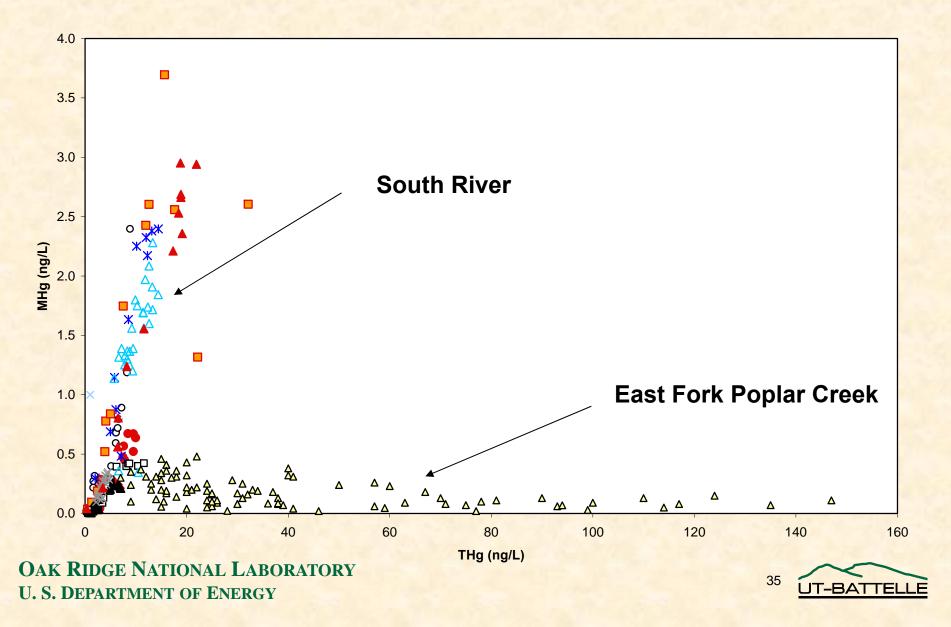


Downstream profiles of methylmercury concentrations (unfiltered) in EFPC, 2000 - 2006, winter versus summer





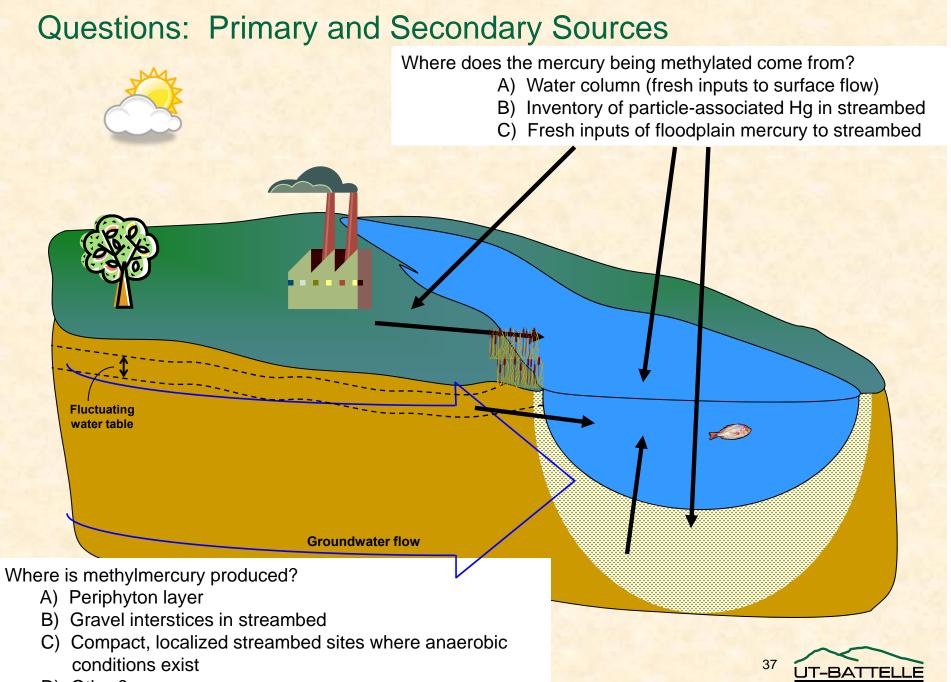
MHg vs THg Dissolved



Important issues regarding bioavailability

- Mercury concentrations in fish throughout EFPC are low relative to the total concentration of Hg in water and sediment. (Low bioavailability)
- Success of Hg remediation efforts requires that bioavailability of Hg in EFPC remain low
- Does increase in Hg bioaccumulation in lower EFPC portend a system-wide change in Hg bioavailability?





D) Other?

Questions: Release Mechanisms

How does MeHg get into the water column?

- A) Resuspension of sediments?
- B) Advection/diffusion from periphyton?
- C) Advection/diffusion from gravel?

Where does inventory of Hg in LEFPC come from, and how fast is it replaced/removed?

Fluctuating water table

- A) What is the inventory of Hg in LEFPC sediments? On periphyton?
- B) What is input rate from floodplain soil? How? (Bank erosion? Larger areas?)
- C) What is annual flux of Hg from EFPC to Poplar Creek? From Y-12 to LEFPC?
- D) Are there depositional hotspots where Hg(0) in streambed inputs?

How do stormflow and baseflow Hg transport interact? A) delayed transit of particle associated Hg through lowermost EFPC?

Questions: Mercury chemical/biological processes; Other factors

What mercury is being methylated?

- A) Dissolved mercury from the N/S Pipe input that never becomes particle- associated?
- B) Dissolved hg in equilibrium with particle-associated hg in water column.

Fluctuating water table

- C) Dissolved mercury desorbed from particulates within the streambed.
- D) Direct methylation of mercury on particles
- E) Hg(0) produced by reduction of Hg(II) in water column or streambed
- F) Reactive mercury produced by oxidation of elemental mercury
- G) Other? (emphemeral Hg(I) species?)

What is rate of MeHg production?

What determines/affects net MeHg production?

A) Factors that affect methylation

B) Factors that affect demethylation

What is the nature of the association of Hg with solids?

Exchangeable (described by Kd)? Biologically incorporated? Precipitate (HgS)? Different in stream than soil?

OAK RIDGE NATIONAL LABORATORY U. S. DEPARTMENT OF ENERGY Groundwater flow

Comparison of South R., Oak Ridge Similarities

- Fish species
- Major ion water chemistry
- Watershed land use
- Degree of Hg-particle association, suspended sediments
- MeHg in fish



Comparison of South R., Oak Ridge Differences

- Source location headwater point source EFPC
 non-point watershed source SR
- Hg source chemistry dissolved EFPC terrestrial soils SR
- Lability of Hg in floodplain soil SR >> EFPC
- MeHg vs HgT positive relationship, SR

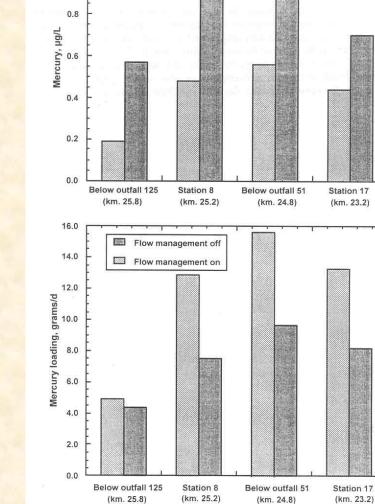
- inverse relationship, EFPC

- Concentration HgT higher in EFPC
- Trace substances Cd, Ni, Cu, Ag, U, PCBs, Zn, Mo elevated in EFPC
- Nutrients NO₃, PO₄ high in EFPC



Shutting off Flow Management temporarily produced a decrease in Hg flux, but an increase in Hg concentration

OAK RIDGE NATIONAL LABORATORY U. S. DEPARTMENT OF ENERGY



Flow management off

1.2

1.0

Fig. 3.17. Effects of raw water shutoff on total mercury concentrations and fluxes at sites in EFPC.



Mass flux of mercury and suspended solids at various sites in EFPC, Dec 19, 2008

